



EFFLUENT DECONTAMINATION USING AQUATIC MOSSSES: A PROMISING APPROACH FOR INDUSTRIAL TREATMENT

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Article History: Received: 10.04.2023

Revised: 25.05.2023

Accepted: 11.07.2023

Abstract

In the present study, the performance of the aquatic moss *Fontinalis antipyretica* for removing cadmium and lead from simulated wastewaters has been evaluated. Five kinetic models were fitted to the experimental data and compared. Previously, the effect of parameters such as the initial solution pH, contact time, and initial metal ion concentration on biosorption was investigated using a sorbent dose of 2 g l⁻¹. The initial pH of the solution was found to have an optimum value in the range of 4.0-6.0. The equilibrium sorption capacity of cadmium and lead by *Fontinalis antipyretica* increased with the initial metal concentration. For an initial metal concentration of 10 mg l⁻¹, the uptake capacity of the moss, at equilibrium, is the same for both metals (4.8 mg g⁻¹). Nevertheless, when the initial concentration increases up to 100 mg l⁻¹, the uptake of Pb(II) is higher than 78%. The pseudo-second order biosorption kinetics provided the better correlation with the experimental data (R² ≥ 0.999).

Keywords: aquatic moss, biosorption, industrial effluent, wastewater treatment.

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DOI: 10.31838/ecb/2023.12.s2.381

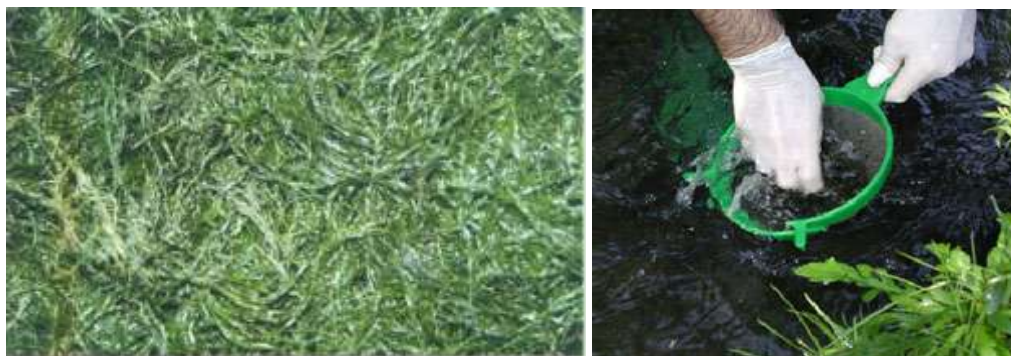


Figure 2 Aquatic moss *Fontinalis antipyretica*.

Batch equilibrium sorption experiments (Fig. 3) were performed in duplicate in 100 mL Erlenmeyer flasks containing 50 mL of metal-containing

solution of a known initial concentration into which a pre-weighed amount of dry biomass was added.



Figure 3 Set of flasks containing metal ions (Cd^{2+} , Pb^{2+}) and biosorbent.

The suspensions were mildly agitated for 300 min on a rotary shaking machine at 140 rpm. Other operating conditions: initial metal ion concentration, 10 and 100 mg L^{-1} ; moss dosage, 2 g L^{-1} ; contact time, 24 h; temperature, 20°C. Kinetic experiments were performed in duplicate at a constant temperature of 20°C, metal concentrations of 10 and 100 mg L^{-1} . Taking into account the results of preliminary experiments, the initial working pH was adjusted in the range 5.0-5.2 by adding 0.1 M H_2SO_4 or 0.1 M NaOH , as required. At fixed time intervals (3, 6, 10, 20, 30, 60, 90, 120, 180 and 300 min) one flask was taken out. Then, the solution was filtered, and concentrations of metal were measured using the Atomic Absorption technique using acetylene-air flame (AAS, VARIAN SPECTRA, model S220). Cadmium and lead standard solutions (1000 $\mu\text{g mL}^{-1}$) were obtained from Merck.

3. Results and Conclusion

Several parameters related to the selected metal, the aqueous medium and the material characteristics could influence the metal removal by mosses. Thus, parameters such as contact time, moss dosage, initial metal concentration, temperature, pH, concentration of Ca^{2+} and nature of the metal salt were studied for the effective

removal of Cd(II) and Pb(II) by the aquatic moss *Fontinalis antipyretica*; however, some results are not showed in this paper. From Cd^{2+} and Pb^{2+} sorption studies, it was evident that pH values equal or higher 4.0 led to higher metal uptake for both metals. At pH 3.0, some metal uptake (20% for Cd and 40% for Pb); according to [12] this may be due to the presence of sulphonate groups that are dissociated around this pH value. This relation of dependence between pH solution and biosorption, suggest that aquatic mosses can be developed as multi-use sorption material. As observed, metal removal is greatly reduced at lower pH, which promote an easy release of metal ions by simple pH adjustment. Therefore, in the subsequent work, experiments were carried out at pH 5.0 - 5.2. Preliminary kinetic tests were carried out at pH 5.0, until maximum metal uptake was achieved, and metal precipitation was avoided. The initial metal concentration in solution was 100 mg L^{-1} , and the moss dosage 2 g L^{-1} . Results evidencing the cadmium and lead sorption dependence on the contact time are plotted in Fig 4. Curves show a similar shape for both metals, characterized by a strong increase in removal capacity by mosses during the first 10 minutes and a subsequent slower phase whose contribution to the total metal uptake is relatively small.

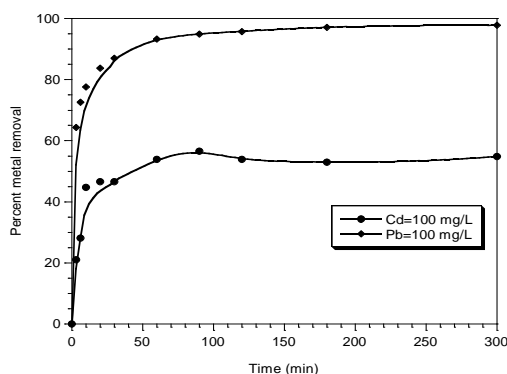


Figure 4 Effect of contact time on Cd(II) and Pb(II) sorption by *Fontinalis antipyretica* ($C_0 = 100 \text{ mg L}^{-1}$; $X = 2 \text{ g L}^{-1}$; $\text{pH} = 5.0$).

At the initial stage was observed a faster removal of ions due to a driving force higher and higher affinity active sites are first occupied. After that, metal concentration decreases and remaining active sites with lower affinities, which are occupied slowly. The sorption process was fast, and the equilibrium attained in less than 2 h. The short time required to attain the equilibrium is a good indicator of the moss potential for a fast removal of metal ions in solution, and it will allow using smaller reactor volumes.

Effect of the initial metal concentration

For 300 min of contact time, 50 mL of metal solution with concentration in the range 10 – 100 mg L^{-1} was shaken with 100 mg biosorbent at constant temperature (20°C) and pH (5.0 - 5.2). Table 1 shows the results of the sorption capacity and rate constants. The first order reaction rate constant, k , was calculated by fitting to the experimental data the equation $C_e = C_0 \exp(-kt)$ for $t \leq 300 \text{ min}$ ($C_e =$ metal concentration at equilibrium, mg L^{-1} and

$C_0 =$ initial metal concentration, mg L^{-1}). The initial metal concentration provides an important driving force to overcome all mass transfer resistances involving the interface liquid/solid, and leads to high metal removal. The equilibrium lead uptake values at different initial metal ion concentrations are given in Table 1 and it is clear that practically all lead ($\approx 98\%$) is bound to the biosorbent, regardless the initial concentration in solution. The affinity of the biosorbent to cadmium is lower and the uptake percentage markedly decreases (from 83.8 to 47.4%) when the initial metal concentration in solution increases from 10 to 100 mg L^{-1} . At lower concentrations, the ratio of initial number of metal ions and the available sorption sites is low and subsequently the biosorption percentage becomes independent of initial concentration. For higher concentrations, however, the available sites of biosorption become fewer and consequently the removal of metallic ions depends on the initial concentration. So, the cleansing yield can be improved by diluting the effluent containing high metal ion concentrations.

Table 1 Effect of initial metal ion concentration on equilibrium adsorption yields and uptake rates of Cd(II) and Pb(II) ($T = 20^\circ\text{C}$; $\text{pH} = 5.0 - 5.2$; $X = 2 \text{ g L}^{-1}$).

C_0 mg L^{-1}	Cadmium			Lead		
	q_e mmol g^{-1}	Uptake %	$k \times 10^3$ min^{-1}	q_e mmol g^{-1}	Uptake %	$k \times 10^3$ min^{-1}
10	0.037	83.8	6.1	0.023	96.7	11.4
20	0.070	78.9	5.2	0.047	97.3	12.0
40	0.131	73.4	4.4	0.096	98.9	14.9
60	0.159	59.8	3.0	0.143	99.0	15.5
80	0.202	56.7	2.8	0.191	98.7	14.5
100	0.211	47.4	2.2	0.235	97.2	11.9

The lead uptake rate is almost independent on the initial concentration and averages $13.4 \times 10^{-3} \text{ min}^{-1}$. On the other hand, the biosorption of cadmium is slower and the uptake rate decreases by about one third as the initial concentration increases from 10

to 100 mg L^{-1} . The high removal capacities observed for the lower initial metal concentrations in solution (Table 1) are explained by the finite number of adsorption sites at the particles surface. However, lead uptake is insensitive to this fact due

to its high electronegativity, which increases the affinity to the different functional groups.

Biosorption kinetic models

Five kinetic models: first-order, pseudo-first-order, Elovich, Ritchie second-order and pseudo-second-order were fitted to the experimental data and compared (only show data and plot for the pseudo-second-order model).

The pseudo-second order equation manipulated ($\frac{t}{q_t} = \frac{1}{kq_e^2} + \frac{t}{q_e}$)

Linear plots of t/q_t versus t are shown in Fig. 4 and k and q_e values, obtained from the slopes and intercepts, are presented in Table 2.

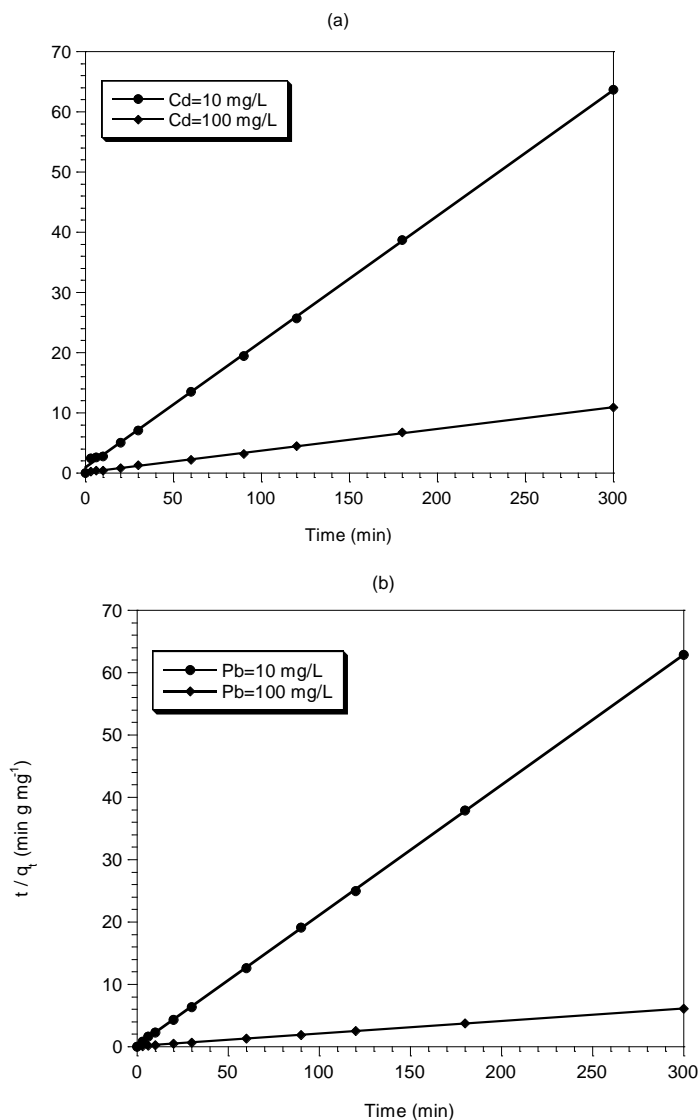


Figure 4 Pseudo second-order sorption kinetics of (a) Cd(II) and (b) Pb(II).

The correlation coefficients (higher than 0.999) indicate a good compliance of the experimental

data with the proposed pseudo-second order kinetic model.

Table 2 Pseudo-second order rate constants for the removal of Cd(II) and Pb(II) by *Fontinalis antipyretica*.

Metal	C_0 mg L ⁻¹	q_e mg g ⁻¹ (mmol g ⁻¹)	k g mg ⁻¹ min ⁻¹	h g mg ⁻¹ min ⁻¹	S^2	R^2
Cd	10	4.80 (0.043)	4.20E-2	0.966	1.49E-1	1.000
	100	27.7 (0.246)	1.00E-2	7.66	1.1E-2	1.000
Pb	10	4.79 (0.023)	0.233	5.34	1.9E-2	1.000
	100	49.3 (0.238)	7.00E-3	16.5	2.6E-4	1.000

The kinetic constant, k , decreased from 4.20×10^{-2} to $1.00 \times 10^{-2} \text{ g mg}^{-1} \text{ min}^{-1}$ as the initial cadmium concentration increased from 10 to 100 mg l^{-1} and a similar behaviour was observed for lead (k decreased from 0.233 to $7.00 \times 10^{-3} \text{ g mg}^{-1} \text{ min}^{-1}$, as the initial lead concentration increased from 10 to 100 mg l^{-1}). The values of the initial sorption rates, h , were determined from the intercepts of the straight lines plotted in Fig. 2. The initial sorption rate increased with the initial metal concentration both for cadmium and lead. This behaviour is opposite to the observed by [13] on the sorption of copper, nickel and lead onto Sphagnum moss peat. Model performance; a brief and simplistic analysis, comparing the regression coefficients (R^2) for the different models of each metal, allows concluded, although generically all values are high (>0.90), the higher correspond to the pseudo-second order model. So, when the models fit the data with sensible values, the suitable procedure is to use an F-test to decide which model to accept. For a level of probability of 95% and 10 degrees of freedom for the numerator and for the denominator, the values of F_{cal} , for each pair and for all models analysed, the pseudo-second order model is statistically more accurate than the others models for the biosorption of Cd(II) and Pb(II) by the aquatic moss, except for cadmium at lower concentration (10 mg L^{-1}), where the Lagergren model is better. However, the Lagergren model is not statistically more accurate than Ritchie model, a second order model ($F_{\text{cal}} < 2.98$). The equilibrium sorption capacities predicted by the best model, for initial cadmium/lead concentrations of 10 and 100 mg L^{-1} , are $0.041/0.023 \text{ mmol g}^{-1}$ and $0.246/0.238 \text{ mmol g}^{-1}$, respectively. The better fit of the pseudo-second order model therefore indicates that a 1:2 binding stoichiometry applies; that is, one divalent metal ion binds to two monovalent binding active sites. It is perfectly reasonable that a divalent metal would bind to two carboxyl groups as showed by Schiewer and Wong (1999), in experiments with brown algae. However, it isn't possible to conclude that, in all the time range, the biosorption reaction is the rate-limiting step. This conclusion may not be valid because a good model fit does not necessarily indicate the real nature of the rate-limiting step. In many other biosorption cases, diffusion rather than the chemical reaction was the rate-limiting step, at least in the initial instants.

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