



TREATMENT OF ORGANIC POLLUTANT IN WASTEWATER BY USING CERIUM OXIDE NANOPARTICLES

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Abstract

In this report, we fabricate spherical cerium oxide (CeO₂) nanoparticles by simple wet chemical method. Utilizing X-ray diffraction (XRD), the crystallinity was investigated. Hexagonal structure of the unit cell was found in the XRD study. The morphology of the fabricated nanoparticles was investigated using Field emission Scanning Electron Microscope (FESEM). Optical band gap was calculated by using UV-Vis spectra. The partial/complete mineralization of organic dye can be utilized using semiconductor nano-photocatalyst. Here, photocatalytic activity of the fabricated CeO₂ was explored by the degradation of methylene blue (MB) dye using visible light irradiation. Different operational parameters e.g. initial dye concentration, catalyst doses and pH and their effect on photodegradation was also studied.

Keywords: Cerium oxide (CeO₂), Photocatalysis, Photocatalytic degradation, Safranin-O, Methylene blue (MB).

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1. Introduction

Wastewaters from textile, paper, medicine and some other industries contains residual dyes and water contaminants, those are not easily degradable [1]. Dyes are amongst the major water pollutants generally present in effluents of textile, leather, paper, printing, cosmetics, petroleum, plastic, food, paint, rubber, and pharmaceutical industries [2]. Additionally, water polluted with dyes can be a potential threat to aquatic life [3]. Several treatment technologies such as adsorption [4-5], oxidation [6], biological treatment [7], coagulation and flocculation [8], membrane processes [9], chlorination [10], ozonation [11], etc. can be used for treatment of wastewater containing dye pollutants. However, due to their complex structure, dyes are very difficult to decolorize using the available techniques [12]. These process however only transfer dyes from liquid to solid phase and further treatment required for secondary pollutant [13]. The use of semiconductor nano-photocatalyst gives the benefit of lower cost, higher efficiency and also suitable recycling [14-15]. As a result, a number of progressive oxidation techniques have been used industrially to speed up the oxidation of

refractory contaminants by producing extremely reactive hydroxyl radicals (OH) [16-17].

Nanoscale cerium oxide (CeO₂) is a wide and direct band gap (3.19 eV) semiconductor [18]. In nanoscale CeO₂ shows tuning band gap and optical properties. They are good candidate for catalytic applications due to their larger exciton binding energy[19-20]. In fact, CeO₂ is one of the most studied catalytic materials, with several examples of its use in accelerating various organic processes in the literature, including carbon monoxide oxidation, catalytic incineration of aromatic hydrocarbons, and mineralization (ozonation) of aniline. Several researchers have reported on CeO₂'s photocatalytic activity [21] and its capacity to degrade synthetic dyes more efficiently than titanium oxide.

2. Materials and methods

Materials

Cerium nitrate hexahydrate [Ce(NO₃)₃.6H₂O] and potassium carbonate [K₂CO₃] with purity ≥99.5% were purchased from Sigmae Aldrich and used as received. Methylene Blue (MB) as a model pollutant from textile industry was obtained from SRL, India. In the photodegradation experiment, a stock solution of MB (1.0 g/L) was used and

diluted to the required concentration with deionized water. All chemical reagents were of analytical grade and used as received. Samples were then preserved in the desiccators over anhydrous CaCl_2 for further use.

2.1. Synthesis CeO_2 nanoparticles

We adopted the conventional wet chemical methods to synthesize zinc oxide nanoparticles (NPs) at room temperature. To synthesize CeO_2 NPs, at first in a beaker 0.05M, 100mL solution of Cerium nitrate was prepared by dissolving $\text{Ce}(\text{NO}_3)_3 \cdot 6\text{H}_2\text{O}$ in distilled water. Similarly, 0.03M, 100mL solution of potassium carbonate (K_2CO_3) was prepared by dissolving K_2CO_3 in distilled water. Aqueous solution of potassium carbonate (50 mL) was added drop by drop with cerium nitrate under constant stirring about 15 h at room temperature. A white precipitate of cerium (III) carbonate was seen in the mixed solution. The precipitation was collected using filter paper and washed with distilled water several times. Resulting precipitation dried in a hot air oven at 60°C for 3 h, followed by calcination at 600°C for 6 h to form the crystalline pure phase.

2.2. Instrumentation

X-ray diffraction (XRD) data were collected in a Rigaku X-ray diffractometer using Cu-K α radiation with the wavelength of 1.54 Å over an angular selection of $20^\circ < 2\theta < 80^\circ$. Optical absorption measurements were accomplished using a Systronics AU 2703 double beam UV-Visible spectrophotometer with a wavelength range of 200-800 nm. At 5 kV accelerating voltages, the surface morphology of the catalysts was investigated using a ZEISS field emission scanning electron microscope (FESEM). The elemental distribution of ZnO was investigated using energy-dispersive X-ray spectroscopy (EDX) on a Leo1430VP microscope with a 5 kV operating voltage.

2.3. Photocatalytic dye degradation experiments

In a cylindrical pyrex-glass cell of 10 cm in height, aqueous solutions containing the photocatalyst were used to degrade methylene blue by photocatalysis. [22]. The suspensions were irradiated under atmospheric condition with white light. The suspension was sampled out at regular intervals during the irradiation. The dispersion was magnetically stirred for 30 minutes (in the dark) prior to illumination to establish adsorption equilibrium between the photocatalyst surface and the methylene blue dye. As a result, the effect of surface adsorption on methylene blue

removal is eliminated. The influence of the pH of the initial solution was evaluated at pHs ranging from 2 to 11 (adjusted using HCl (0.1M) and NaOH (0.1 M)), while methylene blue dye concentration was preset at 30 mg/L of photocatalyst. Photocatalyst dosage was varied from 0.1 to 1 g/L, for 30 mg/L of methylene blue solution, at pH 7. The effect of methylene blue dye concentration was also investigated by adjusting it from 10 mg/L to 80 mg/L at a constant pH of 7 in the presence of 0.7 g/L photocatalyst. Each sample was taken out at regular intervals and rapidly centrifuged at 15,000 rpm for 10 minutes to remove any suspended photocatalysts for analysis. Finally, a UV visible spectrophotometer was used to measure the absorbance of methylene blue in the supernatant liquid at the maximum adsorption wavelength of methylene blue, $\lambda_{max} = 660 \text{ nm}$.

The degradation rate (D) of the dye was calculated using Eq. (1):

$$D\% = (A_0 - A_t/A_0) \times 100 \dots\dots\dots (1)$$

where A_0 represents the initial absorbance of methylene blue solution (blank), and A_t its absorbance after t min of irradiation/reaction.

According to the Beer-Lambert's law A_0 and A_t are proportional to C_0 and C_t , where C_t and C_0 are the concentration of blank and sample at (t) time.

The photodegradation rate constant (k , min^{-1}) was calculated from the slope of the straight-line segment of the plot of $\ln\left(\frac{C_t}{C_0}\right)$ vs. t as a function of the used experimental parameters. The photodegradation of pollutants follows the pseudo first-order kinetics according to the Langmuir-Hinshelwood model [23-24], so the photo decolorization rate of each sample was studied using Eq. (2):

$$\ln\left(\frac{C_t}{C_0}\right) = kt \dots\dots\dots (2)$$

The optical properties of the photocatalysts were studied by UV visible absorption spectroscopy at room temperature.

3. Results and discussion

3.1. Absorption Spectrum

The amount of the energy band gap of CeO_2 NPs was estimated by analysing their optical characteristics using UV visible spectroscopy. The

band gap energy E_g was calculated depending on the absorption spectrum of the sample according to the following Eq. (3):

$$E_g = \frac{1240}{\lambda_g} \dots\dots\dots(3)$$

Where, E_g is the optical band gap energy of the photocatalyst, λ_g is the wavelength in nm used as the absorption edge.

The optical absorption is studied with in wavelength range 200nm to 1000 nm. The observed optical absorption spectrum for the variation of optical absorbance of the grown CeO₂ NPs dispersed in water is shown in Fig. 1(A). The maximum absorbance peak was seen at ≈ 352 nm. The optical absorption coefficient (α) has been calculated from the Lambert law in the wavelength region 200–900 nm. The optical band gap (E_g) of the as-synthesized nanoparticles have been figured utilizing the Tauc equation [25]:

$$(\alpha h\nu)^{\frac{1}{n}} = B(h\nu - E_g) \dots\dots\dots(4)$$

Where α is the absorption coefficient, h is Planck's constant and ν is the frequency ($\nu = \frac{c}{\lambda}$, λ is the wavelength, c is the light speed). For the allowed direct band gap, the value of n equal to 1/2. Exponent n depends on the category of transition and it may have values 1/2, 2, 3/2 and 3 corresponding to the allowed direct, allowed indirect, forbidden direct and forbidden indirect transitions, respectively [26]. B is a constant and

generally called band tailing parameter. Thus, the band gap energy was obtained graphically from $(\alpha h\nu)^2$ vs. $h\nu$ for direct transition, extrapolating the linear part on the abscissa according to Eq. (4). According to this method, the band gap of the samples can be approximated as shown in Fig. 1(B). The estimated direct band gap is 3.52 eV.

Wavelength reliance of optical absorption property of semiconductor can be communicated by the following Eq. (5) [27]:

$$\alpha = \alpha_0 \exp(h\nu/E_U) \dots\dots\dots(5)$$

By plotting $\ln(\alpha)$ with the incident photon energy $h\nu$, the Urbach energy E_U can be resolved. The variation of $\ln(\alpha)$ with the incident photon energy $h\nu$ appeared in Fig. 1(C). Urbach energy is acquired from the inverse of the incline (see Fig. 1(C)). The calculated value of the Urbach energy is 1.026 eV.

3.2. Scanning electron microscope (SEM) analysis

The surface morphology, size, shape and growth mechanism of CeO₂ nanoparticles were investigated through SEM analysis (Fig. 1(D)). These SEM images display closely packed nearly spherical and randomly oriented nanoparticles. There appears some enhancement in the porosity and reduction in the density due to liberation of gases which created surface pores and holes as obvious from these images.

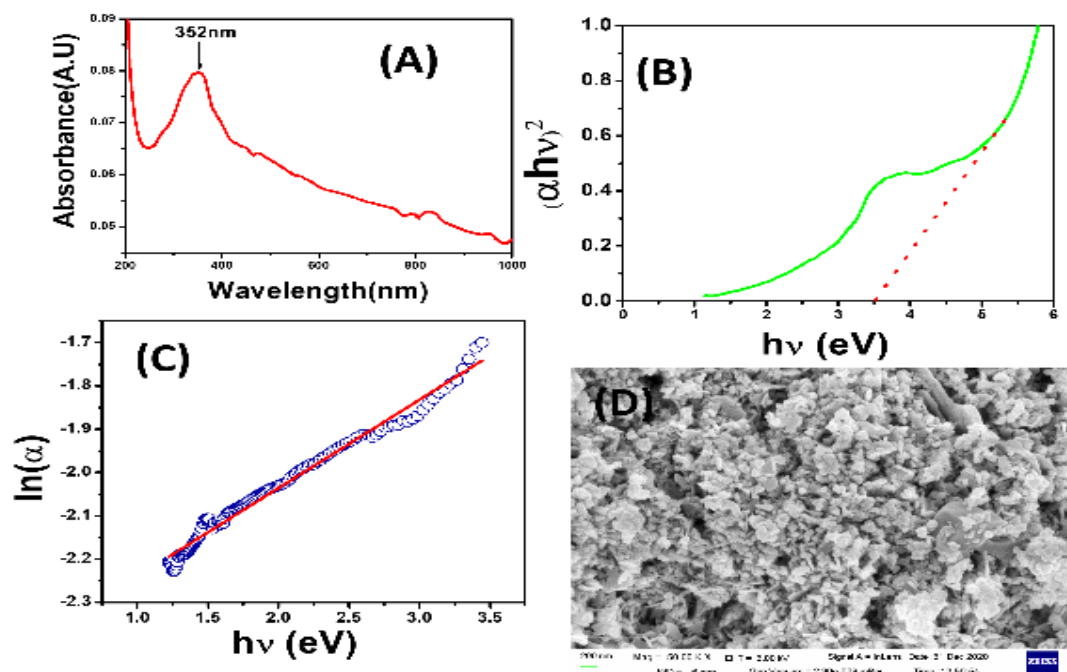


Fig. 1-(A) UV-VIS spectrum of CeO₂ NPs, (B) Tauc's plot for bandgap determination, (C) Plot of $\ln(\alpha)$ vs. photon energy ($h\nu$), (D) FESEM image of CeO₂ NPs

3.3. X ray diffraction

The XRD pattern shown in Fig. 2(A) shows that the unit cell of the synthesized CeO₂ nano-crystal is cubic with the presence of the peaks (111), (200), (220), (311), (222), (400), (331), and (420) [28]. The nanoparticle size is calculated from the Scherrer formula [29]:

$$D = \frac{K\lambda}{\beta \cos\theta} \dots \dots \dots (6)$$

Here, D is the crystalline size and K = 0.94, for copper radiation. The average particle size from this equation is found to be ~ 32.73 nm. The strain (ε) and grain size (D) of the nanomaterial is also

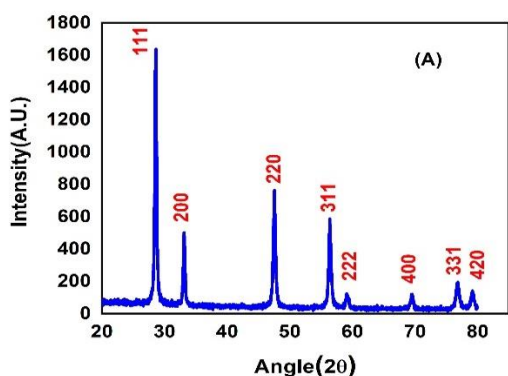
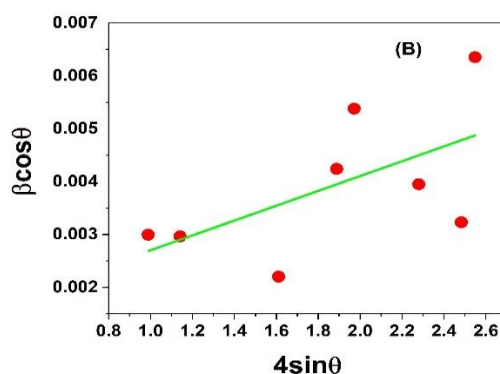


Fig. 2-(A) XRD of CeO₂ NPs, (B) plot of βcosθ vs sinθ



3.4. Effect of operating parameters on the photocatalytic degradation of MB

3.4.1. Effect of catalyst dosage

The rate of photodegradation of a chemical pollutant was found to be a function of photo catalytic dosage in heterogeneous photo catalysis [32-33]. In present work, photo catalyst dosage was varied from 0.1 to 1 g/L, for 30 mg/L of methylene blue solution, at pH 7.

The photocatalytic degradation rates increased as the catalyst load increased. With 0.1 g/L catalyst, the percentage degradation was found to be 38%. When the catalyst load was raised to 0.2 g/L and 0.5 g/L, the percentage was 46% and 59%, respectively, whilst a catalyst load of 0.7 g/L resulted in nearly 87% degradation after 60-70 minutes. Any increase in catalyst load more than 0.7 g/L resulted in poor degradation as shown in Fig. 3(D).

The above observations can be attributed to the availability of active sites on the catalyst surface and the penetration of light into the solution. With increasing catalyst loading, the increased turbidity blocks the light transmission into the solution. The

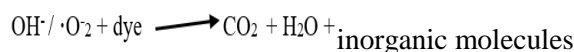
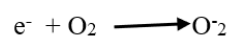
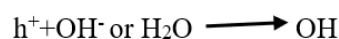
calculated using the Williamson-Hall (W-H) equation [30]:

$$\beta \cos\theta = \frac{K\lambda}{D} + 4\epsilon \sin\theta \dots \dots \dots (7)$$

The average size (D) and dislocation density $d = \frac{1}{D^2}$ are calculated [31]. The average crystal size of CeO₂ NPs is found to be 37.88 nm from the Scherrer formula and 32.73 nm from the fitted result of the W-H equation (Fig. 2(B)). The value of strain is 0.0014 from the fitted result of the W-H equation and 0.0052 from Scherrer equation. The value of dislocation density is $3.127 \times 10^{-3} / \text{nm}^2$ from the result of the W-H and $1.928 \times 10^{-3} / \text{nm}^2$ from the Scherrer equation.

phenomenon is reported in the literature as “scattering” or “screening” effect [32-38].

The degradation can be explained as follows: When photon energy $E > E_g$ [band gap energy] then electrons (e⁻) in conduction band and holes (h⁺) in valance band are generated. The photo-generated holes could either directly oxidize the dye or react with hydroxyl (OH⁻) or H₂O to generate hydroxyl radicals (•OH). The photoelectrons reduce oxygen (O₂) adsorbed on the photocatalyst surface into superoxide radical (•O₂⁻). Finally, dye was decomposed by the generated •OH and O₂⁻. The reaction is given below:



Based on the above observations, 0.7g/L was selected as an optimum catalyst load for further experiments. The photodegradation of MB dye

with catalyst concentration 0.7g/L is shown in Fig. 3(A) and the plot of $\frac{C_t}{C_0}$ vs time is shown in Fig.

3(B). The rate constant saw calculated from Fig. 3(C) and was found 0.035 min^{-1} .

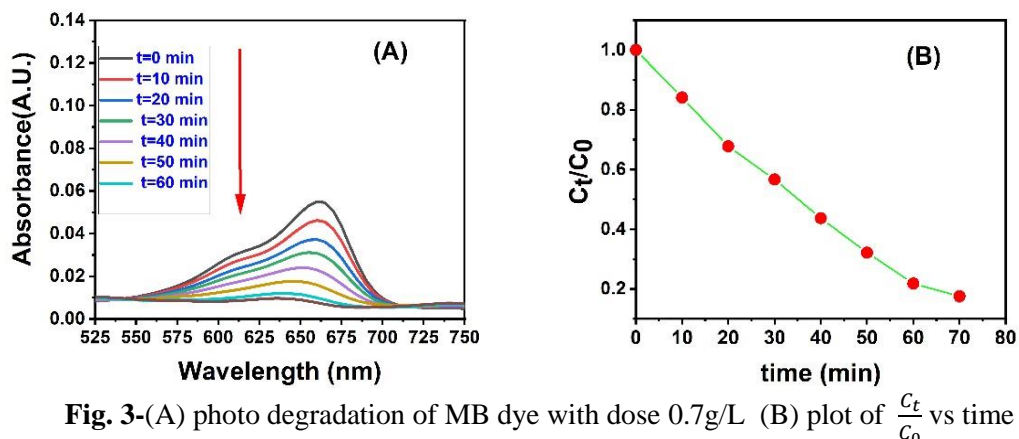


Fig. 3-(A) photo degradation of MB dye with dose 0.7g/L (B) plot of $\frac{C_t}{C_0}$ vs time

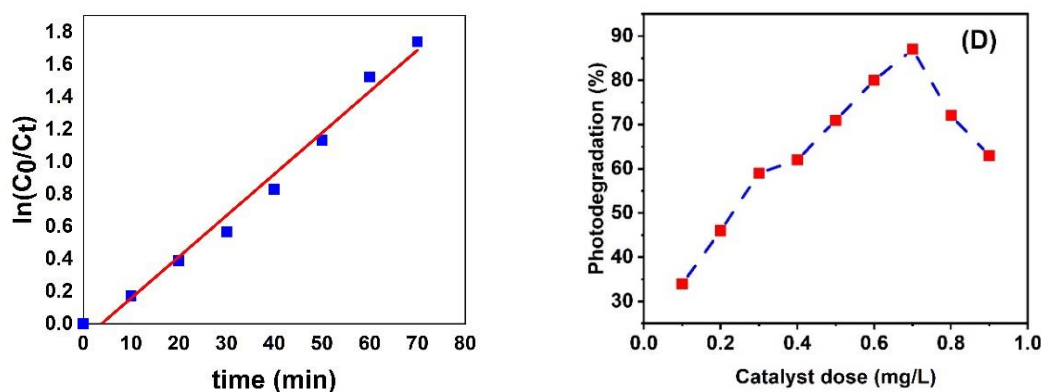


Fig. 3- (C) Photocatalytic degradation kinetic curves (D) Effect of catalyst loading on photocatalytic degradation of MB dye

3.4.2. Effect of initial dye concentration

Initial dye concentration is an indirect measure of opacity. In this study, the effect of initial dye concentration in the range of 10mg/L to 50mg/L was studied at optimum catalyst loading of 0.7 g/L. The results are shown in Fig. 3(E). The variation in the initial dye concentration was selected based on reported in the literature for degradation of various other dyes [39-41]. It is observed that the % dye remaining in the solution increases with increase in the initial dye concentration. For the concentration of 10mg/L, nearly 95% degradation was observed within 60–70 min. However, at an initial dye concentration of 20mg/L, 30mg/L, 40mg/L, 50mg/L the % dye degradation was observed to be 63% and 35%, 26% and 11% respectively. The above observations clearly indicate that photo catalysis reactions are very sensitive to the initial dye concentration.

A comparable pattern has been observed previously in the photodegradation of azo dyes such as Eosin Yellow [42] and Sunset Yellow [43]. The phenomenon is attributed to the formation of $\cdot\text{OH}$ radicals and their probability to

react with dye molecules. When the parameters such as catalyst loading and light intensity are constant, the amount of free radicals generated remains constant. Therefore, at higher initial dye concentrations, an insufficient amount of radicals to target the large amount of dye molecules results in poor degradation.

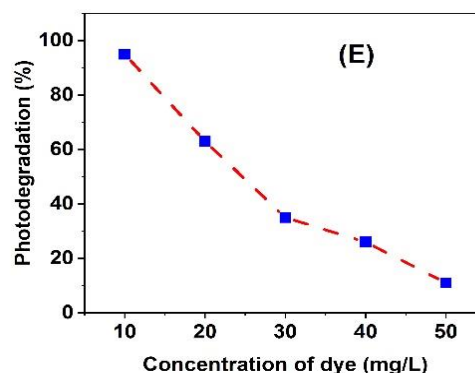


Fig. 3-(E) Effect of dosage on the properties of CeO_2 photocatalysts (F)

3.4.2. Effect of pH

The effect of solution pH in the range of 2–11 on the photocatalytic degradation of 10mg/L MB dye at an optimum catalyst loading of 0.7 g/L was studied. It was seen that at a solution pH of 2, the % degradation of the dye was 85%. Whereas, a nearly complete dye degradation was observed with the solution pH of 3. Increasing the pH value from 3 to 6 (which is natural pH of the solution), did not improved the degradation and the results were comparable to those obtained with pH of 3. At a higher solution pH of 11, the % dye degradation significantly decreased. It is believed that changing the pH of the reaction solution changes the adsorption of the target compound on the surface of the catalyst. The adsorption is expected to be maximum around the isoelectric point (IEP) of the catalyst [44].

4. Conclusion

The present study investigates the potential of chemically synthesized CeO₂ for photocatalytic degradation of methylene blue dye. Complete degradation of the 10mg/L dye solution was achieved within 60 min with 0.7 g/L synthesized CeO₂, natural pH of the solution and at room temperature. The catalyst showed better photocatalytic activity compared to combustion synthesized and commercial catalysts at same reaction conditions. Thus, CeO₂ nanoparticles can be used for waste water treatment and for better environment.

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