



Influence of nutrient enrichment chemistry on species richness and IVI index of aquatic macrophytes: substantiation by Diversity Indices

Rakhi Chahar¹, Rana Mukherji², Sukalyan Chakraborty³, Ravneet Chug¹, Manishita Das Mukherji¹ *

¹Amity institute of Biotechnology, Amity University Rajasthan, Jaipur, India- 303002

²ICFAI Tech, The ICFAI University, Jaipur, India- 302031

³Department of Civil and Environmental Engineering, BIT, Mesra, Jharkhand, India- 835215

*Corresponding author Mail ID: mdmukherji@jpr.amity.edu; manishita@outlook.com

Abstract

The water quality of Bandh Baretha an important freshwater dam is getting impacted by increase in nutrient content due anthropogenic activities conducted nearby. However, studies have not been conducted on how the nutrient chemistry is impacting the water body and the influence of the same on the divergent macrophyte population of this water body. For ascertaining this influence both water and macrophyte samples were collected seasonally from July 2021- April 2022, from three selected sites of the dam. The macrophytes importance value index (IVI), Carlson's Trophic State Index (CTSI), Pearson's Correlation Matrix and Diversity indices were calculated to substantiate the findings. The study found that as nutrients in the lake increased, so did the growth of invasive macrophytes, whose leaves frequently covered the lake's surface leading to degradation of water quality. The increase in the nutrient content is alarming and the authorities and local bodies must work out measures limiting the degradation of the water body.

1. Introduction

Fresh water resources have always been essential for maintaining life and building civilizations. Freshwater reservoirs throughout the world are in danger due to a combination of new and enduring challenges. Water quality is reduced as a result of population explosion leading to enhanced land use practices which pollute waterways with poisonous compounds and excessive nutrients [1,2,3]. A number of investigations have been conducted on the physico-chemical quality of various water bodies to estimate the pollution level and nutrient load in wetland ecosystems throughout the world [4-13] The vital role of macrophytes in the water column is extensive because they play a significant role to the ecosystem's structure, function, and variety, maintain the food chain, and provide habitat for invertebrates and fish [14,15]. Eutrophication is recognized as a type of pollution of water brought on by humans that occasionally results in an aquatic environmental catastrophe and pose threat to the health of the human race and the animal population alike [16,17]. Four stages have been identified for the

various phases and their trophic characteristics: oligotrophic/ultra-oligotrophic, mesotrophic or meso-eutrophic, eutrophic, and hyper-eutrophic [18]. The large accumulation of phosphorus in the water bodies, which is supplied by agricultural land, promotes the eutrophication process [19 - 23].

It is essential to understand the contribution of nutrient chemistry on the seasonal fluctuations in macrophyte diversity and density and ascertain the trophic status of water bodies. With its location in the western portion of the Aravalli mountains and characteristics of a rolling plateau with rounded hills and forest, Eastern Rajasthan is a significant physiographic unit in the state of Rajasthan [24]. The study was conducted on a Dam situated in Eastern Rajasthan to determine the nutrient chemistry of the dam water which influence the macrophyte density and diversity. Further, according to their maximal biomass, the species richness of submerged, floating, and rooted macrophytes was calculated. Along with evenness and dominance indices, the important value index (IVI) of the macrophytes calculated.

2. Materials and methods

2.1. Study area

The three sampling locations were selected at the Dam known as Bandh Baretha. The Dam is situated on the Kakund River and is spread across 36 kms. The region exhibits semi-arid climatic conditions. The nutrient chemistry is influenced by anthropogenic activities which is mainly agriculture. The nutrient content is mainly due to runoff from the neighboring agricultural fields.

2.2. Sampling and analysis

2.2.1. Water sample collection and analysis

The water samples from the three sampling locations were collected in July, October, and December 2021 and April 2022. Water samples were collected at the depth of 0.3-0.5 m. Water temperature (°C), Secchi depth and Dissolved Oxygen (DO) were measured on site. 2 liters of water from each sites was mixed with 2% nitric acid stored and carried to lab at 4°C. All physico-chemical parameters of water were analyzed according to standard methods described by American Public Health Association 2012 [25].

2.2.2. Macrophytes sampling, analysis and diversity of macrophytes

Sampling was done four times for a period of July 2021 to April 2022. Quantitative and analysis of diversity of macrophytes was carried at the three sampling stations. Macrophytes were identified by using dominant taxonomic characteristics. The macrophytes present near the shoreline were handpicked and those deep in water were collected by boat. The macrophytes of the waterbody were identified by visual estimation [26- 30]. Floating quadrant of 1m × 1m was used for counting the number of species of macrophytes and number of macrophytes. For the estimation of dominant species of macrophytes importance value index (IVI) used. IVI was

calculated for the each individual species after calculating relative density, frequency and dominance. Following formulas were used for the calculations:

$$\text{Importance value} = \text{Relative frequency} + \text{Relative density} + \text{Relative dominance}$$

Where,

Relative frequency = number of occurrences of one genus as a percentage of the total number of occurrences of all genera,

Relative density = number of individuals of one genus as a percentage of the total number of individuals of all genera,

Relative dominance = Total area coverage (by visual estimation) of one genus as a percentage of the total area coverage of all genera.

2.3. Carlson's Trophic State Index (CTSI)

A widely used method for describing the trophic condition or health status of a lake is Carlson's Trophic State Index (CTSI). Chlorophyll a (Chl-a), Secchi disc depth (SD), and total phosphorus (TP) are the three variables that Carlson's trophic status index primarily uses to measure algal biomass. Secchi disk was used to measure the secchi depth, disk has 20 cm diameter. The reading was recorded at the deepest point in the lake where the disk was clearly visible when it was lowered. The lakes are categorized as oligotrophic (low productive), mesotrophic (moderately productive), and eutrophic (very productive) based on the values of the Carlson trophic state index. The Table.1 shows the range of the Carlson's trophic state index values and categorization of water.

The formula used for the calculation of CTSI [31] given below:

$$\text{TSI (Chl-a)} = 9.81 \ln \text{Chl-a } (\mu\text{g/l}) + 30.6$$

$$\text{TSI (SD)} = 60 - 14.41 \ln \text{SD (m)}$$

$$\text{TSI (TP)} = 14.42 \ln \text{TP } (\mu\text{g/l}) + 4.15$$

Where TSI = Trophic state index

ln = Natural logarithm

SD = Secchi depth

TP = Total phosphorus

Carlson's trophic state index (CTSI) = [TSI(Chl-a) + TSI(SD) + TSI(TP)]/3.

Table 1 Classification of water on the basis of trophic state index [32].

S.no.	CTSI value	Trophic state	Characteristics of trophic state
1	TSI <30	Oligotrophic	Clear water containing oxygen throughout of the year in upper basin
2	TSI 30-40	Mesotrophic	Water continues to grow aquatic macrophytes, which can cause hypoxia in some shallow water and cause some lakes to become eutrophic in the summer.
3	TSI 40-50	Mesotrophic	Water is moderately clean, but the risk of oxygen depletion rises throughout the summer, and the presence of macrophytes and algae can cause choking condition.
4	TSI 50-60	Eutrophic	Lower classical eutrophic limit, lower transparency
5	TSI 60-70	Eutrophic	Due to excessive growth of macrophytes, blue-green algae and algal scum level of DO decrease
6	TSI 70-80	Hyper eutrophic	Excessive growth of algal bloom throughout of the summer and cyanotoxins algae may be present.
7	TSI >80	Hyper eutrophic	With increasing cyanotoxin algae growth of algal scum become dominant which increases fish mortality, and macrophytes become less in number

2.4. Data processing and analysis

The Carlson's Trophic state index (CTSI) of Dam water was calculated on the basis of Chl-a, SD, TP [33]. Pearson correlation analysis also computed among the physico-chemical parameters of Dam water. The statistical mean with standard deviation of all physico-chemical attributes of water sampled from all three sampling locations of the Dam (table 2).

Table 2: Physico-chemical parameters

Months Parameters	July 2021			October 2021			December 2021			April 2022		
	S1	S2	S3	S1	S2	S3	S1	S2	S3	S1	S2	S3
Temp	30.6 7±0.58	29.67 ±0.8	29.33 ±0.8	25±1.00	24.33 ±0.8	24.67 ±0.8	10.3 3±0.85	10.67 ±0.8	10±1.00	26±1.00	26.33 ±0.8	27.67 ±0.8
pH	6.2±0.1	6.13 ±0.0	6.17 ±0.0	7.3±0.1	7.1±0.09	6.87 ±0.0	7.57 ±0.0	7.37 ±0.0	7.33 ±0.0	6.87 ±0.0	6.77 ±0.1	6.63 ±0.1

	6	6		6	6	6	6	6	6	5	5	
SD	0.31 ±0.0 1	1.37 ±0.0 1	1.53 ±0.0 2	0.46 ±0.0 2	1.36 ±0.0 1	1.62 ±0.0 1	0.73 ±0.0 1	1.32 ±0.0 3	1.78 ±0.0 1	0.96 ±0.0 2	1.45 ±0.0 2	1.83 ±0.0 2
TD	820.	455±	410±	797±	534±	567±	761.	307.6	391.3	478.3	527±	447±
S	67±1 0.12	4.58	9.17	2.65	3.61	4.36	33± 1.53	7±2. 52	3±3. 21	3±12 .58	45.04	27.73
EC	1560 .33±	1469 ±17.	1462 ±37.	1473. 67±4	1357. 33±1	1321. 33±1	127 3±2	1118. 33±2	1238. 67±2	1455. 67±4	1493 ±55.	1392. 67±5
DO	10.2 1.87 ±0.0 6	78 5.2± 00	59 5.5± 0.1	2.15 2.4± 0.1	0.79 6.87 ±0.0 6	7.04 6.73 ±0.1 5	2.61 3.47 ±0.0 6	4.01 10.17 ±0.1 5	6.02 11.2 ±0.2 6	1.53 4.67 ±0.0 6	67 12±0 .5	.13 9.7± 0.26
BO	109. 33±0 .58	75±0 0	75±1 .00	102± 1.00	68±1 .00	67.33 ±2.0 8	97± 2.00	62.33 ±2.0 8	69.67 ±0.5 8	59.33 ±1.1 5	70.67 ±0.5 8	50±1 .00
TH	231± 1.00	305± 5.00	173.6 7±3. 21	163.6 7±3. 51	177.6 7±2. 52	162± 2.65	222. 33± 3.21	162.6 7±2. 52	106.6 7±3. 21	459± 2.65	414± 4.58	339± 4.58
Alk	1372 ±10. 82	1319. 67±4 4.00	1350. 67±3 5.80	1057. 67±3 4.12	1167. 67±1 6.44	1138. 33±2 4.79	103 9±6. 56	1279. 67±1 3.06	1232. 33±4 5.54	1174. 33±1 2.66	1174. 67±2 6.41	1171. 67±1 6.56
Nitr ate	413± 2.65	363.3 3±5. 69	318.3 3±3. 05	382.3 3±2. 09	219± 6.24	202.3 3±1. 15	369 ±1.0 0	222.6 7±1. 53	221.6 7±3. 79	237.3 7±3. 07	243.2 ±2.7 1	247.0 7±5. 18
Nitr ite	201. 67±1 .15	154.3 3±2. 08	132.3 3±1. 53	134.3 3±2. 08	105.3 3±0. 58	113.6 7±2. 08	237. 33± 2.08	157± 1.00	120± 7.94	109.6 7±1. 53	165.6 7±2. 08	120.3 3±0. 58
TP	31.4 7±0. 02	20.01 ±0.0 1	16.41 ±0.0 4	26.03 ±0.0 2	19.28 ±0.0 3	15.09 ±0.0 1	29.1 5±0. 04	19.28 ±0.0 7	13.6 ±0.0 1	24.94 ±0.0 5	17.47 ±0.0 2	12.35 ±0.0 1
Chl -a	6.21 ±0.0 2	5.02 ±0.0 1	3.81 ±0.0 1	5.33 ±0.0 6	3.65 ±0.0 1	3.2± 0.01	6.03 ±0.0 2	4.96 ±0.1 2	3.00 ±0.0 1	4.98 ±0.0 1	3.82 ±0.1 5	2.93 ±0.0 5

3. Results and discussion

3.1. Main physico-chemical characteristics of dam water along with its trophic state

3.1.1. Stie 1: The water quality at site 1st differed noticeably from site 2nd in that site three's pH, TDS, and EC were comparably lower than site 1st, but site 3rd DO was comparatively greater. Higher chlorophyll a content at site 3rd indicates that algal biomass is higher at site 3rd. According to CTSI's assessment, the site's overall trophic condition is eutrophic (Table 3, Fig.1). *Eichhornia crassipes*, as well as rooted and emergent macrophytes, completely covered the site, which may help with phytoremediation by first removing nutrients from the site [34]. This could have assisted in reducing the lake's trophic status from eutrophic to mesotrophic. Aquatic

macrophytes that are invasive and have nutrient enrichment outside of their natural region, like *Eichhornia crassipes*, spread swiftly when introduced under those conditions. This species' exceptional potential for rapid growth has led to its domination in the aquatic ecosystem's trophic structure [35].

3.1.2. Site 2: Site 2nd showed low pH, TDS, DO as compared to permissible limit. Anthropogenic activities from different sources of pollution affect the water quality of Dam. CTSI of the site 2nd was found to be mesotrophic shown in table 4 and Fig.2.

3.1.3. Site 3: Site 3rd is undisturbed area therefore its water quality was better than other two sites. CTSI of this site was found to be mesotrophic as shown in table 5 and Fig.3.

Table 3 Trophic state index of Site 1st

Months	Chl-a (µg/l)	TSI	SD (m)	TSI	TP (µg/l)	TSI
J 21	6.21	48.51	0.31	76.88	31.48	53.88
	6.19	48.48	0.32	76.42	31.45	53.87
	6.22	48.53	0.30	77.35	31.49	53.88
O 21	5.30	46.96	0.45	71.51	26.05	51.16
	5.30	49.96	0.46	71.19	26.02	51.13
	5.40	47.14	0.48	70.58	26.01	51.12
D 21	6.05	48.26	0.73	64.53	29.18	52.79
	6.01	48.19	0.72	64.73	29.11	52.75
	6.02	48.21	0.73	64.53	29.16	52.78
A 22	4.97	46.32	0.98	60.29	24.90	50.51
	5.00	46.39	0.95	60.74	25.00	50.56
	4.98	46.35	0.96	60.59	24.92	50.51
Average		47.77		68.28		52.07

$$\begin{aligned}
 \text{CTSI}_{\text{site 1st}} &= [\text{TSI (Chl-a)} + \text{TSI (SD)} + \text{TSI (TP)}]/3 \\
 &= [47.77 + 68.28 + 52.07]/3 \\
 &= 56.04 \text{ (Eutrophic)}
 \end{aligned}$$

Fig. 1 Three main variables of trophic state index of site 1st

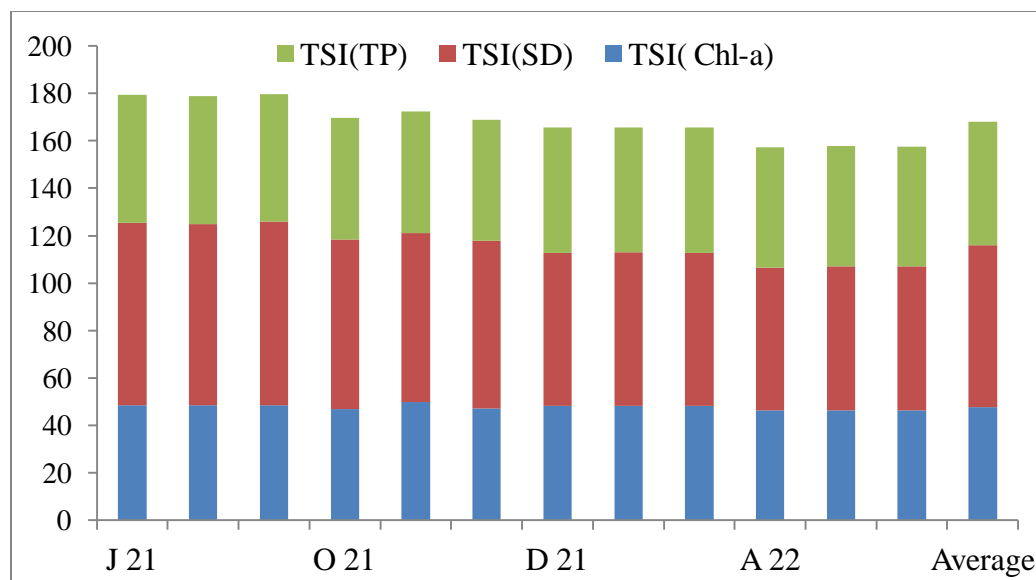


Table 4 Trophic state index of site 2nd

Months	Chl-a (µg/l)	TSI	SD (m)	TSI	TP (µg/l)	TSI
J 21	5.01	46.41	1.36	55.57	20.01	47.36
	5.03	46.45	1.38	55.36	20.01	47.36
	5.01	46.41	1.36	55.57	20.02	47.36
O 21	3.65	43.30	1.36	55.57	19.29	46.82
	3.64	43.27	1.36	55.57	19.25	46.79
	3.65	43.30	1.35	55.68	19.31	46.84
D 21	4.89	46.17	1.3	56.22	19.27	46.81
	4.88	46.15	1.31	56.11	19.36	46.88
	5.1	46.58	1.35	55.68	19.22	46.77
A 22	3.65	43.301	1.43	54.85	17.47	45.39
	3.94	44.05	1.45	54.65	17.45	45.38
	3.86	43.85	1.46	54.55	17.49	45.41
Average		44.93		55.44		46.60

$$\begin{aligned}
 \text{CTSI}_{\text{site 2nd}} &= [\text{TSI (Chl-a)} + \text{TSI (SD)} + \text{TSI (TP)}] / 3 \\
 &= [44.93 + 55.44 + 46.60] / 3 \\
 &= 48.99 \text{ (Mesotrophic)}
 \end{aligned}$$

Fig.2 Three main variables of trophic state index of site 2nd

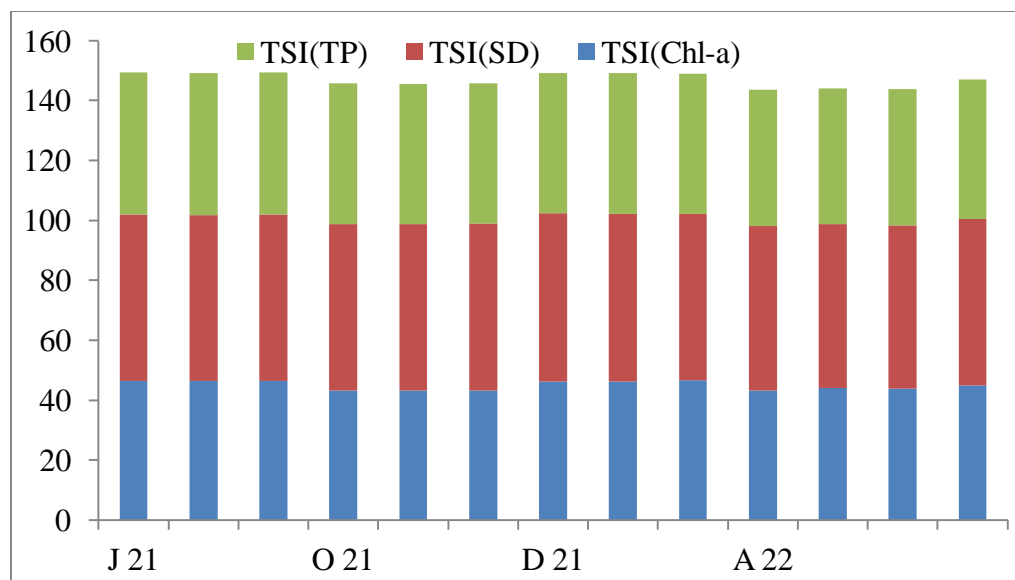


Table 5 Trophic state index of site 3rd

Months	Chl-a (µg/l)	TSI	SD (m)	TSI	TP (µg/l)	TSI
J 21	3.82	43.75	1.55	53.68	16.46	44.54
	3.81	43.72	1.52	53.97	16.38	44.47
	3.81	43.72	1.51	54.06	16.39	44.478
O 21	3.19	41.98	1.61	53.14	15.1	43.30
	3.21	42.04	1.63	52.96	15.09	43.29
	3.2	42.01	1.62	53.05	15.09	43.29
D 21	3.01	41.41	1.78	51.69	13.6	41.79
	3	41.38	1.77	51.77	13.61	41.80
	3	41.38	1.79	51.61	13.59	41.78
A 22	2.99	41.34	1.85	51.14	12.35	40.40
	2.91	41.08	1.81	51.45	12.34	40.39
	2.9	41.04	1.83	51.29	12.36	40.41
Average		42.07		52.48		42.49

$$\begin{aligned}
 \text{CTSI}_{\text{site 2nd}} &= [\text{TSI (Chl-a)} + \text{TSI (SD)} + \text{TSI (TP)}] / 3 \\
 &= [42.07 + 52.48 + 42.49] / 3 \\
 &= 45.68 \text{ (Mesotrophic)}
 \end{aligned}$$

Fig.3 Three main variables of trophic state index of site 3rd

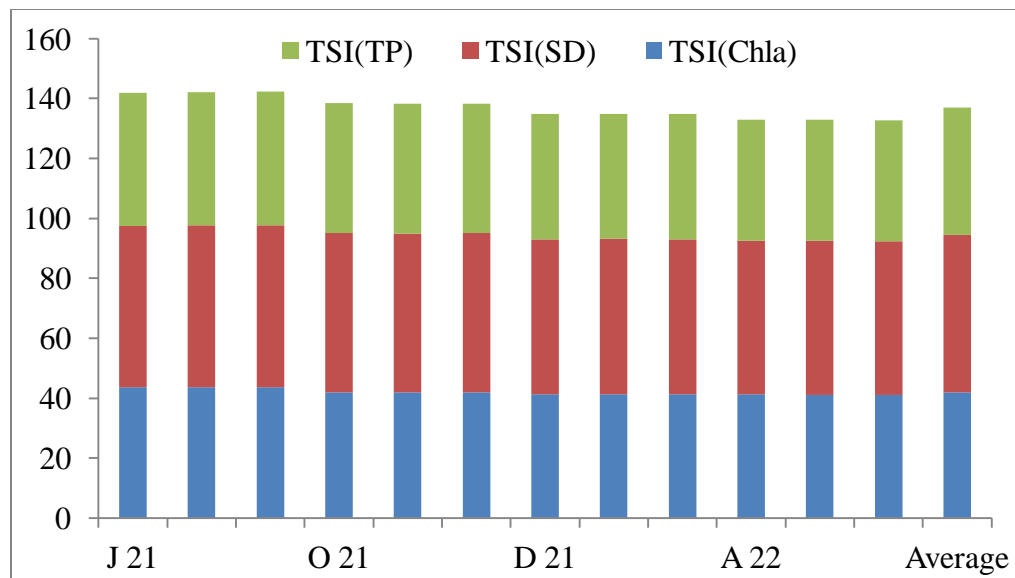
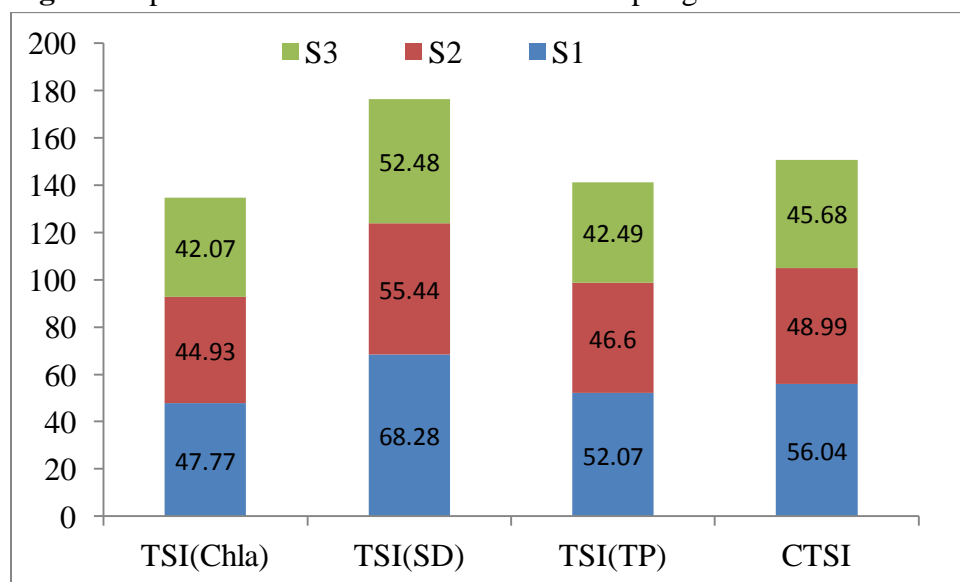


Fig.4 Comparative TSI of the selected three sampling sites



3.2. Statistical analysis in the physico-chemical attributes of selected sites

The five physico-chemical parameters of the dam water, SD, TP, Chl-a, DO, and BOD, responsible for determination of trophic condition, were determined using Pearson's correlation coefficients (Table 6).

At site 1, the correlation coefficients between SD and TP, Chl-a, and DO were all significantly inversely linked with each other. SD and BOD have an inverse correlation. Significantly positive correlations between TP and Chl-a, negative correlations between TP and SD and DO, and positive correlations between TP and BOD were all found.

Site 2nd showed the positive correlation between SD, DO and BOD. TP showed positive correlation with Chl-a and BOD. However, SD showed negatively correlation with TP and Chl-a. TP also showed the negative correlation with DO. Chl-a negatively correlated with DO and BOD. And DO also show negatively correlation with BOD.

Site 3rd showed positively correlation of SD with DO. TP with BOD and Chl-a. Chl-a also positively correlated with BOD. SD negative correlated with TP, Chl-a and BOD. TP was negatively correlated with DO. DO also show negatively correlation with TP and BOD.

Table 6 Correlation Matrix for different sampling sites

		SD	TP	Chla	DO	BOD
SD	Pearson Correlation	1	-.658 [*]	-.623 [*]	.994 ^{**}	-.206
	Sig. (2-tailed)		.020	.030	.000	.520
	N	12	12	12	12	12
TP	Pearson Correlation	-.658 [*]	1	.975 ^{**}	-.657 [*]	.057
	Sig. (2-tailed)	.020		.000	.020	.861
	N	12	12	12	12	12
Chla	Pearson Correlation	-.623 [*]	.975 ^{**}	1	-.640 [*]	.153
	Sig. (2-tailed)	.030	.000		.025	.635
	N	12	12	12	12	12
DO	Pearson Correlation	.994 ^{**}	-.657 [*]	-.640 [*]	1	-.264
	Sig. (2-tailed)	.000	.020	.025		.406
	N	12	12	12	12	12
BOD	Pearson Correlation	-.206	.057	.153	-.264	1
	Sig. (2-tailed)	.520	.861	.635	.406	
	N	12	12	12	12	12

Correlation at Site 2

		SD	TP	Chla	DO	BOD
SD	Pearson Correlation	1	-.771**	-.496	.435	.531
	Sig. (2-tailed)		.003	.101	.158	.075
	N	12	12	12	12	12
TP	Pearson Correlation	-.771**	1	.610*	-.862**	.075
	Sig. (2-tailed)	.003		.035	.000	.816
	N	12	12	12	12	12
Chla	Pearson Correlation	-.496	.610*	1	-.274	-.011
	Sig. (2-tailed)	.101	.035		.390	.972
	N	12	12	12	12	12
DO	Pearson Correlation	.435	-.862**	-.274	1	-.465
	Sig. (2-tailed)	.158	.000	.390		.128
	N	12	12	12	12	12
BOD	Pearson Correlation	.531	.075	-.011	-.465	1
	Sig. (2-tailed)	.075	.816	.972	.128	
	N	12	12	12	12	12

Correlation at Site 3

		SD	TP	Chla	DO	BOD
SD	Pearson Correlation	1	-.982**	-.908**	.915**	-.755**
	Sig. (2-tailed)		.000	.000	.000	.004
	N	12	12	12	12	12

TP	Pearson Correlation	-.982**	1	.914**	-.854**	.846**
	Sig. (2-tailed)	.000		.000	.000	.001
	N	12	12	12	12	12
Chla	Pearson Correlation	-.908**	.914**	1	-.841**	.696*
	Sig. (2-tailed)	.000	.000		.001	.012
	N	12	12	12	12	12
DO	Pearson Correlation	.915**	-.854**	-.841**	1	-.452
	Sig. (2-tailed)	.000	.000	.001		.141
	N	12	12	12	12	12
BOD	Pearson Correlation	-.755**	.846**	.696*	-.452	1
	Sig. (2-tailed)	.004	.001	.012	.141	
	N	12	12	12	12	12

** . Correlation is significant at the 0.01 level (2-tailed).

* . Correlation is significant at the 0.05 level (2-tailed).

3.3. Macrophytes of Baretha Dam

At the study area, 11 macrophytes were observed which included 4 submerged, 5 floating and 2 rooted emergent macrophytes. The dominant macrophyte at site 1st was *Eichhornia crassipes*, and at site 2nd *Hydrilla verticillata* and at site 3rd *Najas minor* was dominant macrophyte species. The dominance was calculated by importance value index (IVI). Macrophytes' diversity was maximum at site 3rd (CTSI = 45.68) in comparison to other sites, indicating that the water was moderately clean. Due to anthropogenic activities trophic state of site 1st (CTSI = 56.04) have highest density of *Eichhornia crassipes* which was evidenced by the lower transparency, risk of oxygen depletion [35].

3.4. Macrophytic diversity indices

The density and diversity of aquatic macrophytes are impacted by changes in land use as well as other stresses, such as urban development, eutrophication, macroeconomic activity, and the

spread of exotic invasive species [36,37]. Simpson's diversity index of site 3rd (0.208) was relatively lower than site 1st (0.23) and site 2nd (0.249) as shown in table 7 and Fig. 5. This indicates that macrophytic diversity increases with decreasing anthropogenic activities and trophic state of the dam. Comparison of CTSI of dam water is shown in Fig. 4. The species richness of submerged macrophytes significantly decreased with rising TP, but the species richness of floating-leaved macrophytes significantly increased. A significant change in trophic status was seen as TP increased. The Shannon-Wiener index of macrophytic species ranged from 1.601 to 1.813. The site with the lowest TP, site 3, had the highest Shannon-Wiener index (1.813), whereas the one with the highest TP, site 1, had the lowest Shannon-Wiener index (1.601). Pielou's evenness index of the dam water ranged from 0.208 to 0.249 having maximum Pielou's evenness at site 2nd and minimum at site 3rd. With lower nutrient loading, the species evenness was found to be decrease. Margalef's diversity index ranged from 1.04 to 1.17. Maximum Margalef's diversity was found at site 3rd and minimum at site 2nd. This examination of diversity makes it abundantly evident that human-induced activities and TP, which determines trophic condition and is a key factor, both have an impact on the macrophytic variety of the lake. The several diversity indices calculated for different sampling sites are shown in table 7.

Fig. 5 Showing comparative picture of macrophytic diversity indices of the sampling sites

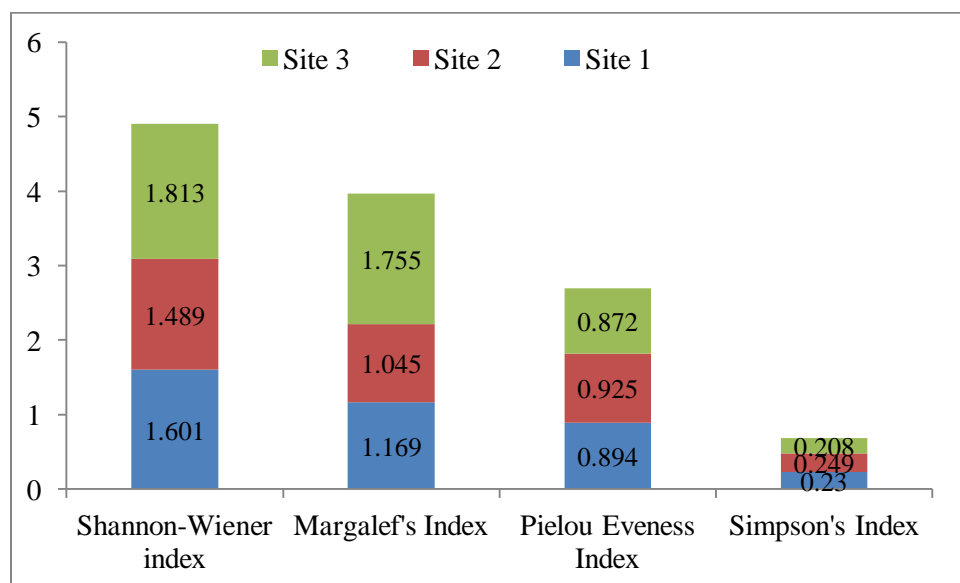


Table 7 Diversity index estimated for different sampling sites

Sampling sites	Site 1	Site 2	Site 3
Shannon-Wiener index	1.601	1.489	1.813
Margalef's Index	1.169	1.045	1.755
Pielou's Evenness Index	0.894	0.925	0.872

Simpson's Index	0.23	0.249	0.208
-----------------	------	-------	-------

Under varied anthropogenic disturbance levels, the Dam exhibits a range of physico-chemical characteristics of water as well as dominance, the composition, and variety of macrophytes. *Eichhornia crassipes*, *Potamogeton crispus*, *Typha angustifolia*, and *Trapa natas* are examples of exotic invasive aquatic weeds that thrive in tropical nations because of the unique eco-hydrological conditions, nutrient-rich runoff, and changes in the watershed land use brought on by human activity [38-41]. The *Eichhornia crassipes* species is the most troublesome free-floating macrophyte whose harmful effects are extensive in aquatic ecosystem. Furthermore, the greater number of exotic invasive species has an effect the native macrophytic diversity of the Dam [35]. The higher nutrient levels have a significant negative impact on submerged macrophytes. Declining natural biodiversity is a result of increased pollution and anthropogenic intervention, as well as the invasion of floating macrophytes.

The species richness of the floated-leaved and submerged macrophytes is significantly affected by total phosphorous (TP), a critical factor determining trophic status [42]. In this study, mesotrophic (TSI = 45.68) site 3rd shows highest species diversity and richness then eutrophic site 1st. According to multiple additional researches, a rise in TP causes an increase in the trophic status of lakes, which leads to eutrophication and decrease in species diversity [43]. Low TP is found to enhance the number of submerged and free-floating macrophytes in lakes, which improves habitat variability and hence species diversity. Submerged macrophytes may occupy a large area in mesotrophic lakes, and their presence has a significant influence on overall structural complexity [44,45].

4. Conclusion

The variations in the hydro-chemical characteristics of Bandh Baretha (Dam) located in Eastern Rajasthan, India was studied. The study revealed the influence of nutrient chemistry on the macrophyte density and diversity. The dominant macrophytes were estimated using the IVI index which were *Eichhornia crassipes*, *Hydrilla verticillata* and *Najas minor*. To have a better understanding of the nutrient influences the trophic status of the water body was also estimated. A Pearson's correlation matrix also substantiated the findings. The various diversity indices also established that the trophic status which is influenced by nutrient content is the driving factor impacting the macrophyte heterogeneity of the study area. The higher nutrient content indicated the degradation of the water body and it is advised that authorities and local bodies must come together to formulate policies for the sustenance of this freshwater ecosystem.

5. References

- [1] Akhtar, N., Syakir Ishak, M. I., Bhawani, S. A., & Umar, K. (2021). Various natural and anthropogenic factors responsible for water quality degradation: A review. *Water (Switzerland)*, 13(19). <https://doi.org/10.3390/w13192660>.

- [2] Kumar, N., Kumar, A., Marwein, B. M., Verma, D. K. I. J., Kumar, A., & Ramamoorthy, D. (2021). Agricultural activities causing water pollution and its mitigation – a review. *International Journal of Modern Agriculture*, 10(1 SE-), 590–609. <http://www.modern-journals.com/index.php/ijma/article/view/632>.
- [3] Zamora-Ledezma, C., Negrete-Bolagay, D., Figueroa, F., Zamora-Ledezma, E., Ni, M., Alexis, F., & Guerrero, V. H. (2021). Heavy metal water pollution: A fresh look about hazards, novel and conventional remediation methods. *Environmental Technology and Innovation*, 22, 101504. <https://doi.org/10.1016/j.eti.2021.101504>.
- [4] Chakravarty, T., & Gupta, S. (2021). Assessment of water quality of a hilly river of south Assam, north east India using water quality index and multivariate statistical analysis. *Environmental Challenges*, 5(August), 100392. <https://doi.org/10.1016/j.envc.2021.100392>.
- [5] Choudhary, S., Sharma, S. K., Sharma, B. K., & Upadhyay, B. (2021). WATER QUALITY ANALYSIS OF ANASAGAR LAKE , AJMER , RAJASTHAN. 11(1), 13–20.
- [6] Jaiswal, M., Hussain, J., Gupta, S. K., Nasr, M., & Nema, A. K. (2019). Comprehensive evaluation of water quality status for entire stretch of Yamuna River, India. *Environmental Monitoring and Assessment*, 191(4). <https://doi.org/10.1007/s10661-019-7312-8>.
- [7] Ju, Y. J., Koh, D. C., Kim, D. H., Mayer, B., & Kwon, H. il. (2023). Evaluating the sources and fate of nitrate in riparian aquifers under agricultural land using in situ-measured noble gases, stable isotopes, and metabolic genes. *Water Research*, 231(December 2022), 119601. <https://doi.org/10.1016/j.watres.2023.119601>.
- [8] Kavindra J, Sharma SK, Sharma BK, & Ojha ML. (2019). Physico-chemical properties and primary productivity of Jawai dam. *Journal of Entomology and Zoology Studies*, 7(4), 865–868. <https://www.researchgate.net/publication/334957797>.
- [9] Pivić, R., Maksimović, J., Dinić, Z., Jaramaz, D., Majstorović, H., Vidojević, D., & Stanojković- sebić, A. (2022). Hydrochemical Assessment of Water Used for Agricultural Soil Irrigation in the Water Area of the Three Morava Rivers in the Republic of Serbia. *Agronomy*, 12(5). <https://doi.org/10.3390/agronomy12051177>.
- [10] Rashid, H., & Prakash, M. M. (2022). Assessment Of Physico Chemical Parameters Of Water And Their Seasonal Variation In Vikram Tearth Sarovar, Ujjain (M.P). *Journal of Pharmaceutical Negative Results*, 13(5), 2102–2110. <https://doi.org/10.47750/pnr.2022.13.S05.332>.
- [11] Sarkar, M., Pal, S. C., & Islam, A. R. M. T. (2022). Groundwater quality assessment for safe drinking water and irrigation purposes in Malda district, Eastern India. *Environmental Earth Sciences*, 81(2), 1–20. <https://doi.org/10.1007/s12665-022-10188-0>.
- [12] Wu, T., Wang, S., Su, B., Wu, H., & Wang, G. (2021). Understanding the water quality change of the Yilong Lake based on comprehensive assessment methods. *Ecological Indicators*, 126(April), 107714. <https://doi.org/10.1016/j.ecolind.2021.107714>.
- [13] Yıldız, S., & Karakuş, C. B. (2020). Estimation of irrigation water quality index with

- development of an optimum model: a case study. *Environment, Development and Sustainability*, 22(5), 4771–4786. <https://doi.org/10.1007/s10668-019-00405-5>.
- [14] Beltran, R., Beca-Carretero, P., Marbà, N., Jiménez, M. A., & Traveset, A. (2020). Spatio-temporal variation in macrofauna community structure in Mediterranean seagrass wrack. *Food Webs*, 25, e00178. <https://doi.org/10.1016/j.fooweb.2020.e00178>.
- [15] Turcios, A. E., Miglio, R., Vela, R., Sánchez, G., Bergier, T., Włodyka-Bergier, A., Cifuentes, J. I., Pignataro, G., Avellan, T., & Papenbrock, J. (2021). From natural habitats to successful application - Role of halophytes in the treatment of saline wastewater in constructed wetlands with a focus on Latin America. *Environmental and Experimental Botany*, 190(July). <https://doi.org/10.1016/j.envexpbot.2021.104583>.
- [16] Badamasi, H., Yaro, M. N., Ibrahim, A., & Bashir, I. A. (2019). Impacts of Phosphates on Water Quality and Aquatic Life. *Chemistry Research Journal*, 4(3), 124–133. <https://chemrj.org/download/vol-4-iss-3-2019/chemrj-2019-04-03-124-133>.
- [17] Ukaogo, P. O., Ewuzie, U., & Onwuka, C. V. (2020). Environmental pollution: Causes, effects, and the remedies. In *Microorganisms for Sustainable Environment and Health*. INC. <https://doi.org/10.1016/B978-0-12-819001-2.00021-8>.
- [18] Bera, B., Bhattacharjee, S., Shit, P. K., Sengupta, N., & Saha, S. (2021). Anthropogenic stress on a Ramsar site, India: Study towards rapid transformation of the health of aquatic environment. *Environmental Challenges*, 4(April), 100158. <https://doi.org/10.1016/j.envc.2021.100158>.
- [19] Bol, R., Gruau, G., Mellander, P. E., Dupas, R., Bechmann, M., Skarabøvik, E., Bieroza, M., Djodjic, F., Glendell, M., Jordan, P., Van der Grift, B., Rode, M., Smolders, E., Verbeeck, M., Gu, S., Klumpp, E., Pohle, I., Fresne, M., & Gascuel-Oudou, C. (2018). Challenges of reducing phosphorus based water eutrophication in the agricultural landscapes of Northwest Europe. *Frontiers in Marine Science*, 5(AUG), 1–16. <https://doi.org/10.3389/fmars.2018.00276>.
- [20] Gross, A., Reichmann, O., Zarka, A., Weiner, T., Be'eri-Shlevin, Y., & Angert, A. (2020). Agricultural sources as major supplies of atmospheric phosphorus to Lake Kinneret. *Atmospheric Environment*, 224(May), 117207. <https://doi.org/10.1016/j.atmosenv.2019.117207>.
- [21] Lin, S. S., Shen, S. L., Zhou, A., & Lyu, H. M. (2021). Assessment and management of lake eutrophication: A case study in Lake Erhai, China. *Science of the Total Environment*, 751, 141618. <https://doi.org/10.1016/j.scitotenv.2020.141618>.
- [22] Ni, Z., Wang, S., Wu, Y., & Pu, J. (2020). Response of phosphorus fractionation in lake sediments to anthropogenic activities in China. *Science of the Total Environment*, 699, 134242. <https://doi.org/10.1016/j.scitotenv.2019.134242>.
- [23] Sarvajayakesavalu, S., Lu, Y., Withers, P. J. A., Pavinato, P. S., Pan, G., & Chareonsudjai, P. (2018). Phosphorus recovery: a need for an integrated approach. *Ecosystem Health and Sustainability*, 4(2), 48–57. <https://doi.org/10.1080/20964129.2018.1460122>.

- [24] Chahar, R., & Mukherji, M. Das. (2022). Impact of Climate Change on Aquatic Plants Found in Major Lentic Water Bodies Located at Eastern Rajasthan. *Ecology, Environment and Conservation*, December, 508–510. <https://doi.org/10.53550/eec.2022.v28i03s.074>.
- [25] APHA (American Public Health Association). (2012). Standard Methods for the Examination of Water and Wastewater. *Standard Methods for the Examination of Water and Wastewater*, 1496.
- [26] Dubey, D., Kumar, S., & Dutta, V. (2022). Impact of nutrient enrichment on habitat heterogeneity and species richness of aquatic macrophytes: evidence from freshwater tropical lakes of Central Ganga Plain, India. *International Journal of Environmental Science and Technology*, 19(6), 5529–5546. <https://doi.org/10.1007/s13762-021-03438-4>.
- [27] Dulenin, A. A., Kharitonov, A. V., & Sviridov, V. V. (2022). Selection of an Optimum Method for the Assessment of the *Saccharina japonica* Stock in the Northwestern Tatar Strait. *Russian Journal of Marine Biology*, 48(7), 678–687. <https://doi.org/10.1134/S1063074022070045>.
- [28] Kislik, C., Genzoli, L., Lyons, A., & Kelly, M. (2020). Application of UAV imagery to detect and quantify submerged filamentous algae and rooted macrophytes in a non-wadeable river. *Remote Sensing*, 12(20), 1–24. <https://doi.org/10.3390/rs12203332>.
- [29] Sharma, R. C. (2017). Macrophytes of Sacred Himalayan Lake Dodi Tal, India: Quantitative and Diversity Analysis. *Biodiversity International Journal*, 1(4). <https://doi.org/10.15406/bij.2017.01.00020>.
- [30] Ghosh, D., & Biswas, J. K. (2015). Biomonitoring Macrophytes Diversity and Abundance for Rating Aquatic Health of an Oxbow Lake ecosystem in Ganga River Basin. *American Journal of Phytomedicine and Clinical Therapeutics*, 3(10), 2321 – 2748.
- [31] Carlson. (1977). *Carlson 's Trophic State Index for the assessment of trophic status of two Lakes in Mandya district*. 3(5), 2992–2996.
- [32] Carlson, R. E., & Simpson, J. (1996). A Coordinator's Guide to Volunteer Lake Monitoring Methods. *North American Lake Management Society*, 96. <https://www.nalms.org/secchidipin/monitoring-methods/trophic-state-equations/>
- [33] Lencha, S. M., Tränckner, J., & Dananto, M. (2021). Assessing the water quality of lake hawassa Ethiopia—Trophic state and suitability for anthropogenic uses—applying common water quality indices. *International Journal of Environmental Research and Public Health*, 18(17). <https://doi.org/10.3390/ijerph18178904>.
- [34] Ting, W. H. T., Tan, I. A. W., Salleh, S. F., & Wahab, N. A. (2018). Application of water hyacinth (*Eichhornia crassipes*) for phytoremediation of ammoniacal nitrogen: A review. *Journal of Water Process Engineering*, 22(October 2017), 239–249. <https://doi.org/10.1016/j.jwpe.2018.02.011>.
- [35] Ajithram, A., Jappes, J. T. W., & Brintha, N. C. (2021). Water hyacinth (*Eichhornia crassipes*) natural composite extraction methods and properties - A review. *Materials*

- Today: *Proceedings*, 45(xxxx), 1626–1632. <https://doi.org/10.1016/j.matpr.2020.08.472>.
- [37] Senetra, A., Dynowski, P., Cieślak, I., & Żróbek-Sokolnik, A. (2020). An evaluation of the impact of hiking tourism on the ecological status of alpine lakes-a case study of the valley of dolina pieciu stawow polskich in the tatra mountains. *Sustainability (Switzerland)*, 12(7). <https://doi.org/10.3390/su12072963>.
- [37] Ziaja, M., Wójcik, T., & Wrzesień, M. (2021). Phytosociological data in assessment of anthropogenic changes in vegetation of rzeszów reservoir. *Sustainability (Switzerland)*, 13(16). <https://doi.org/10.3390/su13169071>.
- [38] Abell, J. M., Özkundakci, D., Hamilton, D. P., & Reeves, P. (2022). Restoring shallow lakes impaired by eutrophication: Approaches, outcomes, and challenges. *Critical Reviews in Environmental Science and Technology*, 52(7), 1199–1246. <https://doi.org/10.1080/10643389.2020.1854564>.
- [39] Asmare, T., Demissie, B., Nigusse, A. G., & GebreKidan, A. (2020). Detecting Spatiotemporal Expansion of Water Hyacinth (*Eichhornia crassipes*) in Lake Tana, Northern Ethiopia. *Journal of the Indian Society of Remote Sensing*, 48(5), 751–764. <https://doi.org/10.1007/s12524-020-01107-6>.
- [40] Kumari, K., Swain, A. A., Kumar, M., & Bauddh, K. (2021). Utilization of *Eichhornia crassipes* biomass for production of biochar and its feasibility in agroecosystems: a review. *Environmental Sustainability*, 4(2), 285–297. <https://doi.org/10.1007/s42398-021-00185-7>.
- [41] Yu, H., Dong, X., Yu, D., Liu, C., & Fan, S. (2019). Effects of Eutrophication and Different Water Levels on Overwintering of *Eichhornia crassipes* at the Northern Margin of Its Distribution in China. *Frontiers in Plant Science*, 10(October), 1–11. <https://doi.org/10.3389/fpls.2019.01261>.
- [42] Yi, R., Song, P., Liu, X., Maruo, M., & Ban, S. (2019). Differences in dissolved phosphate in shallow-lake waters as determined by spectrophotometry and ion chromatography. *Limnology*, 21(3), 329–339. <https://doi.org/10.1007/s10201-019-00574-2>.
- [43] Wang, H., García Molinos, J., Heino, J., Zhang, H., Zhang, P., & Xu, J. (2021). Eutrophication causes invertebrate biodiversity loss and decreases cross-taxon congruence across anthropogenically-disturbed lakes. *Environment International*, 153, 106494. <https://doi.org/10.1016/j.envint.2021.106494>.
- [44] Poikane, S., Kelly, M. G., Várbiro, G., Borics, G., Erős, T., Hellsten, S., Kolada, A., Lukács, B. A., Lyche Solheim, A., Pahissa López, J., Willby, N. J., Wolfram, G., & Phillips, G. (2022). Estimating nutrient thresholds for eutrophication management: Novel insights from understudied lake types. *Science of the Total Environment*, 827. <https://doi.org/10.1016/j.scitotenv.2022.154242>.
- [45] Muthukrishnan, R., & Larkin, D. J. (2020). Invasive species and biotic homogenization in temperate aquatic plant communities. *Global Ecology and Biogeography*, 29(4), 656–667. <https://doi.org/10.1111/geb.13053>.

